Original Research **Properties of Coal Gasification Wastes Essential to Determining Their Impact on the Environment**

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Abstract

Gasification is playing an increasingly significant role in thermal processing. The different types of solid waste generated in these processes should ultimately be treated as by-products and be used in various industries (such as fly ash from coal combustion, for example). However, studies of waste properties are needed to determine real and potential impact on the environment. The scope of our tests depends on methods and directions of planned use of waste. Our paper presents and discusses the test results of waste generated in research installations: one in a large laboratory and the second in a pilot plant. The wastes are derived from the process of direct combustion, direct gasification, and the two-stage process of coal gasification with char formation followed by combustion. All installations are based on the fluidized bed, with two hard coals and lignite used as fuel. Tested materials belong to two groups: fly ash and bottom ash. Research of the waste properties was conducted in the field of grain size distribution, chemical composition, content of combustible leaching of chemical contaminants, and radioactivity. Analysis of the results was carried out both in terms of existing legislation and also related to a comparison of properties of materials from single-stage and two-stage thermal processing.

Keywords: fly ash, gasification waste, leaching, radionuclides, fluidization bed

Introduction

Each thermal processing of solid fuels, including coal, is associated with the generation of waste/solid by-products stemming from the mineral matter contained in the fuel. The most common process of this type is, of course, coal combustion, and solid waste generated in this way is predominantly fly ash. Over the last decades fly ash, and more broadly energy wastes, have stopped being considered mainly as burdensome waste, problematic to disposal and a potential risk to the environment, and have begun to be more widely regarded as a valuable by-product used in many industries. However, to make this possible it was necessary to start conducting in-depth and systematic research to determine the possible adverse impact on the environment.

Adverse impacts of ash may involve a number of elements of the environment, including in particular the atmosphere, water, and soil, as well as live nature, including humans. An important problem may be dusting of particulates like fly ash, both during their manufacture and transport. The risk of water and soil contamination is mainly related to the possibility of leaching of chemical contaminants, including heavy metals [1]. Properties and composition of the eluates depend primarily on the fuel properties of the source [2]. The problem of leaching of chemical contaminants can concern both the same ash, as well as materials and building components containing them, such as embankments, road substructures [3], or other places of use and disposal of such materials. Another important part of the environmental assessment is to determine the possible radiation emission from the ashes, making it necessary to determine the content of radionuclides [4, 5].

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In most cases, the adverse environmental impact of ash can be avoided. It is necessary, however, to have full information about the key parameters of such materials. This concerns information about the kinds and basic physico-chemical properties of the ash, which depends on both the properties of the fuel and the type of thermal processing technology. The last several years have seen an increase in the amount of ash from plants based on fluidized bed boilers. Waste from such installations can be divided into two basic groups: fly ash, in which most of the particles are smaller than 0.1 mm, and bottom ash, where sand fraction predominates [6, 7]. A common feature of both types of ash is alkaline reaction, and the share of flue gas desulfurization process (usually calcium compounds).

Beside combustion, in recent years the gasification process has been playing an increasingly important role in the field of thermal processing of solid fuels [8-10]. So far, Poland does not possess any of this type of installation on an industrial scale. Advanced research and development in this area is being conducted with the use fluidized bed reactors. The direct product of coal gasification in this installation is a char, which then was burned in a fluidized bed boiler.

As a result of these operations and tests, solid wastes have been produced in the form of fly and bottom ash. The analysis properties of this ash in light of the potential environmental risk is the subject of this article.

Materials

The test materials were derived from the thermal processing of three different coals: lignite and two hard coals. The main series of material samples for testing are the result of two thermal processes of coals. In the first stage two types of hard coal and one lignite were gassified at atmospheric pressure in a pilot fluidized bed reactor belonging to the Institute for Chemical Processing of Coal (IChPW) in Zabrze, Poland [11-13]. In this process, semiproducts (char coals) were formed containing carbon in amounts from 40% to about 60%. In the next step char coals were combusted in a fluidized bed boiler (the largelaboratory scale), belonging to the Technical University of Czestochowa (PCz) [14]. Furthermore, in the PCz installation the same coals were directly combusted, from which char coals were prepared. In these two ways, 12 samples of ashes were produced (six bottom and six fly), while six were derived from coal combustion and six from char coal combustion. Designation of samples is as follows: the first letter indicates the type of source coal (B - lignite from Belchatow Mine, J - hard coal from Janina Mine, and W hard coal from Wieczorek Mine), the second, directly combusted material: W - coal (or lignite) and K - char. The third letter indicates the type of ash: L – fly ash and D – bottom ash. For example:

- BKD bottom ash from combustion of char from lignite 'Belchatow' gasification
- WKL fly ash from combustion of char from 'Wieczorek' hard coal gasification
- JWD bottom ash from combustion of 'Janina' hard coal

The second group of materials represents two samples: fly ash (FA3) and bottom ash (BA1) produced during the tests of direct gasification process of lignite in the PCz installation, working then as a fluidized bed gasifier [14].

Coals of which ash formed (directly or indirectly) may be characterized by several important parameters such as ash content in the dry matter (A^a) and the content of: the volatiles (V^{daf}), carbon (C^{daf}), hydrogen (H^{daf}), and sulfur (S^{daf}) in terms of pure coal material. These values for coal B are: A^a=19.9%, V^{daf}=57%, C^{daf}=67.1%, H^{daf}=1.1%, S^{daf}=1.0%, for coal J: A^a=12.8%, V^{daf}=42.1%, C^{daf}=79.6%, H^{daf}=5.5%, and S^{daf}=4.0%, and for coal W: A^a=13.0%, V^{daf}=38.2%, C^{daf}=83.8%, H^{daf}=5.8%, and S^{daf}=1.71% [15, 16].

The Scope and Methods of Wastes Testing

Energy waste impact on the environment is a very broad issue and over the years has been described in a number of publications, also as a compilation and review form, in probably every country where the waste is produced. It is worth noting that this problem is still the subject of evaluations, studies and investigations due to the great diversity of this type of material, as well as a multitude of directions and conditions of use or disposal [6, 17-27]. Based on previous experience in research of waste from the combustion of coal and other waste materials, the physical and chemical properties of waste/by-products of coal gasification were identified, which is important for determining the potential environmental risk. This group includes grain composition, leaching of chemical contaminants, and content of radionuclides, and - to a lesser extent - chemical composition and content of combustible parts.

Grain Composition, Dust Fraction Content

Detailed determination of particle size distribution of waste, part of the basic research necessary for the characteristic of the material, is performed before deciding on the direction of waste management. This test may be useful to assess the possible negative impact of waste on the environment, especially on the atmosphere and living organisms, including humans. This is particularly important in the case of materials with a high proportion of the particulate fraction, less than 63 μ m. The content of the respirable fraction ($<3 \mu m$) is especially important for the health and safety of people and animals. Such grains can enter the lungs of living organisms and accumulate there. Knowing the size distribution of the material, it is possible to determine the source and extent of possible contamination, and undertake appropriate precautions to prevent dust emissions during storage and transport, or specific methods of utilization.

Particle size distribution of waste from coal gasification depends mainly on the location of their formation, processing conditions, and type of installation. The highest content of the particulate (dust) fractions shows the ash captured in gas purification systems, especially those derived from the system based on the fluidized bed. The particle size of the gasification by-products depends mainly on the type of installation where the byproducts are generated. The biggest content of slit fraction is shown in the ash captured in gas purification systems, especially from the installations based on a fluidized bed.

Radioactivity (Radionuclide Content)

The concentration values of natural radioactive isotopes are given according to ITB regulations (Warsaw, No. 234, 1995). The measurements have been conducted with the use of a gamma radiation spectrometer [28]. The methods and equipment must meet the requirements of the Council of Ministers of 2 January 2007 (Journal of Laws, No. 4 Item 29) on the requirement for the natural radioactive isotopes in minerals and materials for civil buildings, industrial waste used in building industry, and the controlled content of these isotopes. According to the regulation mentioned above the values of indicators f_1 and f_2 can go above $f_1=1$ and $f_2=200$ Bq/kg.

The regulation PN-G-11011:1998 describes the maximum content of radionuclides in materials used for backfill in underground mining. According to this the specific activity of radioactive isotopes (radium isotope 226 Ra+ 228 Ra) should not be above 1×10^4 [Bq/kg].

The conducted literature research did not reveal any information on the increased radionuclide content in the waste from coal gasification. Nevertheless, such research for ash both from combustion and gasification are of common practice and inseparable parts of their characteristics for both utilization and landfill [4, 5, 29-33].

Unburned Carbon Content

In gasification by-products sometimes the grains of char (the not totally processed coal) may occur. This may cause inconvenience in the management of these materials, as well as the storage (landfill) and utilization - hence there is a risk of uncontrolled oxidation and thus the emission of gases into the atmosphere. Such problems have appeared in the past mainly in the case of extractive waste in coal mining. Nowadays, high carbon content in an ash is very rare because of the limitation in management of these waste and the effectiveness of fuel use. In the case of systems in industrial scale the occurrence of the higher content of flammable parts in the ash proves the faults of some stage in the process of thermal processing and usually causes immediate preventative actions. Content of unburnt carbon was determined indirectly on the basis of loss on ignition tests at 700 and 1000°C.

Chemical Composition

The chemical composition of the by-product from coal gasification depends on many factors, mainly the chemical composition of the fuel itself but also the additives used (e.g. the ones that increase or decrease the temperature of ash melting), and sorbents or any material from the fluidized bed (in the case of the reactor of this type). Among the energy wastes from the combustion processes available on the market the products of fume desulphurization are of significant meaning in the assessment of environmental hazards. Particularly noteworthy are the results of studies on the content of calcium oxide, sulphur oxide, and heavy metals.

The tests were conducted with use of spectroscopy (ICP-AES), and mass (ICP-MS) emission methods, as well as an energy-dispersive x-ray fluorescence (EDXRF) spectrometer X-5000.

Chemical composition testing is of auxiliary character, while discussing the environmental hazard, allowing for a preliminary assessment of the potential in range of leaching of chemical contamination, and showing the possible direction of their management. This may be required when the specific methods of the waste use is applied. Due to the lack of relevant provisions regarding this kind of waste, content of heavy metals in the ashes compared with their permissible content determined for soils in industrial areas is of significant importance.

The Leaching of Chemical Contamination and Reaction

Tests on the leaching of chemical contamination is the basic type of test that allows us to determine the possible impact of waste on water, soil, and organisms. The objective of these tests is to determine the amount of contamination that can penetrate the environment as a result of water operating. Water is a basic media for contamination by the waste managed both by storage and utilization, including that used in land engineering applications such as heaps or earthwork – either as components of building materials or as objects such as roads, houses, etc. The experience gained so far in Poland and elsewhere underlines the fact that the leaching test is significant for determining the ways and methods of dealing with the products of thermal processing due to the possibility of long-lasting migration of the contamination in the environment [25, 34, 35].

The range of tests covers mainly the content determination of the chlorides and sulphates, ChZT, cyanide, sodium, potassium, and many of the heavy metals such as chrome, zinc, cuprum, nickel, lead, or cadmium. The tests were conducted with use of spectroscopy (ICP-AES) and mass (ICP-MS) emission methods.

The determination of pH is essential both to assess the influence on the surface and underground water, and on living organisms. The increased value of pH (alkaline) may be caused by the use of the additives modifying the properties and that are used to capture sulphates.

The Tests Results

Grain Composition

The potential environmental hazard relating to particle size distribution of the materials is associated primarily with the risk of dust. Therefore, the conducted research concentrated mainly on fly ash. The results of grain compositions of fly ash derived from the char combustion and ash obtained from lignite gasification (FA3) are shown in Fig. 1 and marked by black lines. Grey lines represent comparative materials derived from combustion of hard coal (FFC) and lignite (FFL) in industrial power plants with fluidized bed boilers, as well as fly ash from hard coal combusted in a power plant with a pulverized boiler [36].

It should be underlined that the grain distribution of fly ash from the combustion of hard coal chars (WKL and JKL) is very similar to the fly ash from the coal combustion in commercial power plant with fluidized bed boiler (FFC), but differs from grain size distribution curve of the fly ash from coal combustion in industrial pulverized boiler (FPC). Differences in grain composition of fly ash originating from lignite are more noticeable. The materials derived from research installation (FA3, BKL) shows bigger content of smaller grains than the ashes from commercial power plants – both based on lignite (FFL) and hard coal (FFC).

The content of grains smaller than 63 μ m was, in the studied fly ash above 70%, what shows the serious risk of dustiness in the case of not obeying safety regulations. The content of the respirable dust is also significant: the fraction <3 μ m is between 5% and 15%. The results obtained do not differ greatly from results for typical fly ash from lignite and coal combustion in industrial boilers with a fluidized bed [37].

Radioactivity (Radionuclide Content)

Research has been conducted with the use of a semiconductor gamma radiation spectrometer owned by the Department of Physics and Nuclear Techniques AGH, according to the CoM ordinance of 2 January 2007 on regulating the amount of natural radioactive isotopes in raw materials used for civil and agricultural buildings, and in industrial waste used in the building industry and the controlling of isotope content. This document defines the limit values for parameters f_1 and f_2 as follows: $f_1 \le 1\pm 20\%$ and $f_2 \le 200$ Bq/kg $\pm 20\%$. Radionuclide content in the tested samples does not exceed the allowance limits for building materials (Table 1). In most, the content of the radionuclides in the tested ash was lower than typical for of such materials [5, 6, 30, 38, 39]. However, in the case of fly ash from 'J' coal the results for f_1 reach the allowed limits. In such a situation additional tests should be conducted for different samples taken at intervals.

The Content of Unburned Carbon

Among the tested samples, some fly ash taken from the large laboratory installation showed a significant amount of unburned coal (Table 1). This appears to be due to low effectiveness of the fluidized combustion process for two hard coals and chars from gasification of this coal. It was probably due to a mismatch of process parameters to properties or the fuels. Such situations occur more often in laboratories rather than in industrial installations. Nevertheless, this raises a lot of problems in the use of such waste [40-43]. The content of unburned coal reaching amounts greater than 2% is rather worrying, and content above 8% eliminates the ash from application in building or mining industries, without additional processing.

In the case of bottom ash and fly ash from lignite the amount of unburned coal was at a much lower level.

Chemical Composition

Tests on chemical composition of fly ash are of basic importance for determining its properties while thermal processing, and to indicate the possible sources of increased leaching of chemical contamination. The chemical composition of ash from char combustion (without loss on ignition) is shown in Table 1. The increased value of SiO_2 in the bottom ash (WKD, BKD) indicates the presence of the material from the fluidized bed – in this case quartz sand.

The content of specific heavy metals in the studied material is diverse (Table 1). Therefore, for the evaluation of the results three factors were taken into account whose



Fig. 1. Grain-size distribution of ash from char combustion.

				Limit for	soil in indus-	trial places				500	300	600	1000	60	p/u	30	600											
	FA3	0.66	0.84							1		1	1	1	1	1	1											
	BA2	2.53	4.56																									
	WWL	27.49	34.97		246	53	38	0.45	53	128.7	67	82	147	24.7	20.9	22.3	93.3											
	WWD	0.73	1.2		301	27	26	0.32	27	396	230	28	142.2	6.2	0	5.4	69.2											
	JWL	15.21	20.62		540	152	108	1.23	152	169.7	113.3	98	282.3	47.7	6.1	22.3	181.3											
	JWD	0.62	0.79		420	39	40	0.46	39	136.7	66.7	40.7	260.3	15.7	0	14	241											
	BWL	2.44	3.57		77	93	58	0.63	93	96.3	41	40.3	71	23.3	23.4	19	36.7											
	BWD	0.25	0.4		207	26	22	0.27	26	548	179.7	43	145.7	11.3	0	2.7	69											
	WKL	38.97	45.32	oactivity	228	44	31	0.38	44	106.3	0	66	216.3	24.3	17	23.3	115.3	59.96	7.75	9.26	10.16	1.09	0.79	5.06	5.49	0.22	0.08	0.15
ials.	WKD	1.35	1.84	Radi	324	29	28	0.34	29	137.3	19.3	39	134.3	9.4	0	6.3	44.7	74.41	3.59	10.39	4.81	0.51	0.36	3.58	2.19	0.07	0.02	0.05
sted mater	JKL	29.66	35.26		414	125	83	0.97	125	225.3	60.7	111.3	389	40.3	8.9	27.7	128	45.17	39.35	5.34	2.61	1.90	1.71	1.39	1.28	0.78	0.35	0.12
ntent in te	JKD	0.43	0.66		554	97	69	0.86	97	130.3	17.3	89	228.7	55.7	0	18.7	90.7	49.67	38.97	2.91	4.53	0.48	0.58	0.79	1.75	0.15	0.12	0.06
y metal co	BKL	2.94	9		72	118	73	0.78	118	90.8	8.5	58	113.3	28.8	31.4	28	14.3	50.24	8.10	2.61	30.20	0.16	0.15	0.86	7.43	0.18	0.04	0.03
clide heav	BKD	0.53	0.89		254	55	40	0.46	55	134.3	43	40.3	137	12.7	1.9	10.3	47.7	67.06	13.86	4.34	8.73	0.34	0.32	2.05	3.03	0.14	0.06	0.06
nd radionu	Unit	[%]	[%]		[Bq/kg]	[Bq/kg]	[Bq/kg]	Ξ	[Bq/kg]	[mdd]	[mdd]	[mdd]	[mdd]	[mdd]	[mqq]	[mdd]	[mdd]	[%]	[%]	[%]	[%]	[%]	[%]	[%]	[%]	[%]	[%]	[%]
ignition an	0	700°C	1000°C		40K	²²⁶ Ra	228 Th	f_1	f_2	Cr	Ż	Cu	Zn	As	Se	Hg	Pb	SiO_2	Al_2O_3	Fe_2O_3	CaO	Na_2O	K_2O	MgO	SO_3	P_2O_5	BO_2	TiO_2
Table 1. Loss of	Sample	T acc of imitian	LOSS OI IgIIIIOII			Specific activity		Indiantam	IIIUICAUUIS				Heavy metals	content									Chemical composition	-				

potential impact was recognized to be the most important: the type of coal from which they are derived, the type of ash (fly and bottom), the process of their preparation (direct coal combustion or combustion of char resulting from coal gasification), and combinations thereof.

The highest contents of chromium (Cr) and selenium (Se) were recorded in the ash derived from coal (lignite) Belchatow, copper (Cu), zinc (Zn), arsenic (As), mercury (Hg), and lead (Pb) in ash from the coal 'J,' and nickel (Ni) in the ash from coal 'W.' Studies have found generally higher contents of copper (Cu), selenium (Se), and mercury (Hg) in the fly ash as compared to the bottom ash. However, for all heavy metals a significant role was also played by other factors. For each metal, the situation was as follows:

- In the case of chromium, its content in all ash from coal 'J' was similar, but significantly bigger differences were observed for the remaining ash. Definitely the largest chromium content was noted for bottom ash from combustion of coal samples (BWD and WWD).
- The highest nickel content was recorded in bottom ash. Furthermore, it is generally more noticeable content of this element in samples from the direct combustion of coals compared to the combustion of chars.
- Higher average value of copper (Cu) were found in fly ashes and in all the wastes from the combustion of chars, regardless of the type of coal from which they originated.
- The contents of Zn and As are slightly differentiated in particular types of waste and largely dependent on the origin of the waste (type of coal). However, the arsenic content is generally lower in the bottom ash (except for waste from the combustion of char 'J').
- Selenium is practically non-existent in bottom ash. The highest content of Se was measured in fly ash from hard coals (coals 'J' and 'W').
- Significantly higher mercury content was found in fly ash as compared to bottom ash, as well as in wastes from the combustion of chars.
- Average lead content was greater (for samples from 'B' and 'J' significantly greater) in materials from the combustion of coals, and generally in bottom ash.

Despite the differences claimed only in one case, the copper content in the bottom ash BWD acceptable limits laid down for soils from industrial sites was exceeded. However, the high diversity of the results indicates a need to also consider the type of thermal processing, in addition to the type of coal and ash in future studies of the content of heavy metals in generated wastes.

Leaching of Chemical Contamination

The results of testing on leaching of chemical contamination of ash from coal and char combustion are shown in Table 2. The table contains a column that shows the limit values for eluates of waste used in backfilling technologies in hard coal mining. The requirements are consistent with the regulation of the Ministry of the Environment of 24 July 2006 on the condition that it should be applied while introducing wastewater into the water or soil, and on the substances harmful to the aquatic environment (Polish Journal of Laws 2006 No. 137 Item 984).

The preliminary analysis of the results shows few examples of the values above the allowed limits. This refers to pH and the content of sulphates. Such a situation occurs frequently, especially in the case of ash from coal combustion in commercial fluidized bed installations or ash containing products from fume desulphurization [6, 37, 44]. Such materials are approved for use in the mining industry by separate decisions of the State Mining Authority.

The obtained results are rather differentiated and, to a great extent, depend on the kind of fuel they originate from. They do not allow us to clearly state the influence of the thermal processing on the leaching of chemical contamination. In searching for general tendencies, the direct comparison has been conducted on the same kinds of waste coming from combustion of coal (sample 'W') or char from coal (sample 'K'). Moreover, the analysis has been conducted of the average values calculated basing on dependence on the kind of fuel combusted (coals or chars from gasification), the type of waste (fly ash 'L' or bottom ash 'D'), and both of these indicators simultaneously. The following relations and tendencies have been observed:

- No matter of the type of fuel, significantly higher values of leaching both for cations and anions have been observed for eluates of fly ash in comparison with bottom ash.
- In almost all cases ash 'K' showed lower pH than ash 'W'; the biggest differences refer to pairs JKD-JWD and WKD-WWD originating from hard coals.
- Higher values of cations, mainly heavy metals, have been observed in eluates of 'K' samples. This refers to potassium, cadmium, and copper for all samples, nickel for samples based on hard coals, and mercury in the case of samples from 'J' ash. However, it should be noted that the absolute value of these contaminations stays at a low level.

In other cases direct comparisons do not allow for more general conclusions.

Conclusions

The research has been conducted on waste from the direct coal combustion process in large-laboratory scale installations, combustion of char from coal gasification in a pilot installation, and direct gasification of coals. Both installations are based on fluidized bed technology. Three types of coals were used for testing: two hard coals and lignite.

The main aim of the study was to characterize this type of waste to determine the potential environmental risk in case of their future development and use. Moreover, attention has been paid to the comparison of the properties of ash from coal and char combustion.

Most of the materials meet the requirements on the environmental impact with regard to the two most impor-

Table 2. Values	of the leacl	hing of che	mical conta	umination (c	ontent in w	vater eluate:	s) for ash fi	rom coal ga	sification a	nd burning	char obtain	ed from co	al gasificati	ion.		
Samples / Indicator	JKL	JKD	MKL	WKD	BKL	BKD	JWL	JWD	MWL	DWWD	BWL	BWD	BA1	FA3	Limit	values
Hq	9.19	8.75	9.18	8.37	12.29	9.58	9.62	12.26	11.58	11.68	12.25	10.39	8.05	7.83	; 9	-12
Component								mg/dm ³								
$\mathrm{Na}^{\scriptscriptstyle +}$	71.73	13.09	59.13	7.92	12.01	4.56	79	7.48	34.54	3.53	15.72	5.86	0.20	225.30	800	sodium
$\mathbf{K}^{\scriptscriptstyle +}$	14.83	12.35	19.48	5.9	6.34	3.28	11.25	4.36	8.66	1.18	3.79	2.47	1.49	53.16	80	potassium
Zn^{+2}	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	0.001	2.0	zinc
Cu ⁺²	0.0014	0.0005	0.0008	0.0117	0.0033	0.0103	0.0004	0.0001	0.0003	0.0001	0.0001	0.002	0.0002	0.0001	0.05	copper
Ni^{+2}	0.002	0.0095	0.0015	0.0106	0.001	0.0002	0.001	0.0003	0.0005	0.002	0.002	0.003	0.0001	0.1140	2.0	nickel
Pb^{+2}	0.0002	0.0001	0.0002	0.005	0.0005	0.0006	0.0002	0.002	0.0002	0.001	0.0015	0.0007	0.0007	0.0020	0.05	lead
Hg^{+2}	0.0051	0.0047	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.02	mercury
Cd^{2+}	0.00125	0.0004	0.0005	0.0001	0.0003	0.0002	0.001	0.0002	0.0003	0.00005	0.0002	0.00015	0.0002	0.0004	0.1	cadmium
Cr^{3+}	0.003	0.0034	0.033	0.0075	0.047	0.034	0.002	0.002	0.019	0.0002	0.138	0.018	0.0001	0.0040	0.5	chrome
AS^{3+}	0.0120	0.0480	0.0100	0.0030	0.0210	0.0031	0.0460	0.0160	0.0130	0.0200	0.0170	0.0180	0.0006	0.004	0.2	arsenic
CI ⁻	66	2.9	307	4.9	312.0	16	16.2	1.7	590.0	8.0	89.9	3.9	5.7	233.5	1 000	chlorides
SO_4^{-2}	1028	483.1	798.3	435.6	212.0	420.8	845.1	372.6	987.2	420.3	453.9	345.2	109.10	1361.0	500	sulphates
Total cations	86.6	25.5	78.7	13.9	18.4	7.9	924.5	749.4	1 594.90	450.6	768.3	382.2	85.8	846.9		
Total anions	1127	486	1105.3	440.5	524	436.8	1 524.0	1 174.5	2 394.8	651.0	1 289.4	550.5	5.7	1686.2		

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tant tests: leaching and radionuclide content. This is the basis for the initiation of technological research to determine the specific directions of use of such wastes.

In several cases exceeding the limits were found (especially in terms of pH and sulphate content), which is typical also for some of the energy waste generated in the Polish power plants. These are not completely disqualifying features of that waste from the possibility of its use since there are methods known to reduce environmental risks in this area, particularly in the mining industry. It should be noted, however, that there is virtually no possibility of using fly ash with a high carbon content (and thus high loss on ignition), in the most popular directions as construction or mining, without additional processing.

During the studies large differences were found in the properties of the tested ash, especially in terms of their chemical composition and leaching. Analysis of the results indicates that the type of thermal process in which the ash is formed has a significant influence on their properties, followed by such factors as the type of ash and coal from which it originates.

Our studies did not identify a potential environmental hazard of tested materials significantly larger than from ash generated in coal combustion in power plants. However, exceeding the permissible limits, in some cases, indicates a need for further, regular testing of this type of ash. It is necessary to continue this type of research for both definition of the scope of variability of properties relevant to determine the impact on the environment, as well as further analysis of the observed relationships. Such studies should be performed for different types and process parameters as well as type of fuel.

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